

Monitoring of breeding birds in Umbria, Central Italy, between 2000 and 2005

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Abstract – Common species of birds breeding in Umbria were monitored from 2000 to 2005 by means of an annual repeated survey of 1574 point-counts scattered across the whole region. 20 points were placed in each 10x10 km UTM grid unit, located 1 km apart from each other along low traffic roads, crossing the most representative environments. We obtained yearly population indices for 67 species and examined population trends; in order to avoid possible first-time effects, we performed all the analyses omitting the first year. Furthermore we calculated multi-species indicators for species associated with farmland and woodland habitat, taking the geometric means of annual specific population indices. 42 % of the species for which trends could be produced have declined during the years 2001 to 2005, whilst only 28 % have increased. Multi-species indicators seem to show a better situation for farmland than for woodland birds.

INTRODUCTION

Since the breeding season 2000, the “Osservatorio Faunistico Regionale dell’Umbria”, in cooperation with “Provincia di Perugia” and “Provincia di Terni”, has been carrying out annual surveys to produce yearly population indices for common species of breeding birds. The regional monitoring scheme collaborates with a similar scheme across the rest of Italy (Progetto MITO2000 - Fornasari *et al.*, 2002).

Lately, driven by the need to elaborate effective methods to measure changes in biodiversity, multi-specific indicators for farmland bird species (Farmland Bird Index - FBI) and woodland bird species (Woodland Bird Index - WBI), have been developed and proposed at a European level (Gregory *et al.*, 2005).

Here we describe: a) population trends observed in Umbria over the period of the study; b) the results obtained through calculating FBI and WBI from the data collected.

METHODS

Umbria (Central Italy) is one of the smallest Italian regions (8450 km²). Without coasts, it is characterized by an eastern highland and a western sector with a predominantly hilly morphology. The most widespread habitats are woodland and agricultural cultivations, covering 41,4 %

and 41,5 % of the region respectively; urban and industrial settlements comprise only about 5 % of the area.

During six consecutive breeding seasons (2000-2005) a team of 14 ornithologists carried out an annual survey of 1574 points (the same each year), scattered across the whole region.

Points were selected using the following sampling pattern: in each 10x10 km UTM grid unit, we traced a route along low traffic roads, paying attention that it would pass at least through the most representative environments; along the selected routes, points were placed 1 km apart from each other, assuring for each grid unit a density of one sampling point every 5 km². We verified *a posteriori* that the selected points were broadly representative of habitats within the region, although with a bias towards agricultural cultivations (49,3 % in the sample vs. 41,5 % in the study area), while woods were under-represented (30,7 % vs. 41,4 %).

The method used in the field was a version of point-counts without distance limit from the observer, during a period of 10 minutes each.

For 67 species recorded in at least 1 % of all stations, we used the software package TRIM version 3.53 (Pannekoek & van Strien, 2005) to compute population indices on an annual basis and to assess their trends. TRIM uses a log-linear Poisson regression procedure that gives a trend estimate (**b**). This is the exponential of the slope parameter of a linear regression between years (more exactly the

Table 1. Criteria to classify trend estimates.

| 95% confidence interval of b | | TREND designations |
|-------------------------------------|-------------------|--------------------|
| lower limit | upper limit | |
| > 0.95 and ≤ 1.00 | < 0.95 | steep decline |
| | <1.00 and ≥ 0.95 | moderate decline |
| > 1.00 and ≤ 1.05 | ≥ 1.00 and < 1.05 | stable |
| | | moderate increase |
| > 1.05 | | strong increase |
| ≤ 0.95 | ≥ 1.00 | uncertain |
| ≤ 1.00 | ≥ 1.05 | uncertain |

independent variable is “year - 1”) and the corresponding yearly counts transformed as natural logarithm; **b** expresses the mean annual change of the counts during the period of the study. TRIM supplies a confidence interval (at 95 % probability level) of the trend, on the basis of which the trend is classified (Tab. 1). We used TRIM in the following way: each point has been considered as a location itself; no covariates were inserted; model type 2 (*linear trend*) has been used, with the *stepwise* procedure to select the points where the slope of the log-linear equation changes (*change-points*); to estimate the equation parameters and their standard errors we turned to the GEE (Generalised Estimating Equations) procedure.

Since many species showed a steep increase between the first and the second year of the survey, we suspected that it might depend to a large extent on the increasing surveyors’ experience. Therefore, in order to avoid these very likely startup effects (reported in other monitoring programmes too - Kendall *et al.*, 1996), we omitted year 2000 and restricted all analysis to the period from 2001 to 2005.

Among the 67 common species, 27 species were typical of agricultural habitat and 19 species characteristic of woodland habitat, according to the classification worked out by the MITO2000 project (Progetto MITO2000, 2006). For these two groups, farmland and woodland species, we determined two multi-specific indicators of the population trends, as proposed by Gregory *et al.* (2005). Such indicators are calculated by simply taking the geometric means of annual specific populations’ indices. In doing this, we followed two slightly different approaches:

- a) in one case we applied the methodology shown by Gregory *et al.* (op.cit.). For this indicators are produced for species that are largely associated with farmland or woodland habitat irrespective of the habitat of the sampling points;
- b) in the other case, we repeated the calculation proce-

dures of the specific populations’ indices taking into account the points placed in agricultural habitat only (N = 826) for the farmland species, and the ones placed in woodland habitat only (N = 459) for the woodland species. In this way, the indicators reflect changes within the habitat of interest.

RESULTS

In Tab. 2 we compare the population trend estimates (**b**) computed using all data (i.e. surveys 2000-2005) with the estimates calculated omitting the first year (i.e. surveys 2001-2005): 54 species out of 67 (80,6 %) show a decrease of the trend estimate if the first survey is removed. This result strongly suggests that a startup effect actually occurred and justifies the exclusion of the first year from the analysis.

In Tab. 3 the trends of the 67 species between 2001 and 2005 are given in more detail: 10 species (14,9 %) show a strong increase; 9 (13,4 %) a moderate increase; 19 (28,4 %) a moderate decline; 9 (13,4 %) a strong decline; 5 (7,5 %) are stable and 15 (22,4 %) show an uncertain trend.

The trends of FBI and WBI obtained through the procedure that takes into account the whole set of sampling points are in good concordance with the trends obtained taking into account only the points belonging to the proper environment (Fig. 1). This concordance reflects the high proportion of species (70,4 % amongst farmland and 73,7 % amongst woodland birds) showing a significant correlation between populations’ indices obtained with the two different methods (Tab. 4, Tab. 5); it is worth to note that sometimes the correlation is high, even if only a little part of the sites in which a species was found is included in the habitat of interest (see for example the chaffinch *Fringilla coelebs* and the blackcap *Sylvia atricapilla* in Tab. 5).

WBI decreased substantially whilst FBI showed a de-

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Table 2. Comparison between the trend estimates computed using all data (b_0) and omitting the data from the first year (b_1), with corresponding standard errors. For each species, the difference between the two estimates ($b_1 - b_0$) is shown.

| years | 2000-2005 | | 2001-2005 | | $b_1 - b_0$ | SPECIES | years | 2000-2005 | | 2001-2005 | | $b_1 - b_0$ |
|-------|-----------|-------|-----------|-------|-------------|-------------------------------|-------|-----------|-------|-----------|--------|-------------|
| | b_0 | SE | b_1 | SE | | | | b_0 | SE | b_1 | SE | |
| | 1,007 | 0,024 | 0,929 | 0,026 | -0,078 | <i>Buteo buteo</i> | 1,120 | 0,018 | 1,064 | 0,021 | -0,056 | |
| | 1,075 | 0,030 | 1,075 | 0,039 | -0,001 | <i>Falco tinnunculus</i> | 0,975 | 0,016 | 0,937 | 0,019 | -0,038 | |
| | 1,044 | 0,020 | 1,158 | 0,031 | 0,114 | <i>Coturnix coturnix</i> | 0,910 | 0,025 | 0,858 | 0,030 | -0,052 | |
| | 0,986 | 0,010 | 0,953 | 0,012 | -0,032 | <i>Phasianus colchicus</i> | 1,056 | 0,004 | 0,984 | 0,005 | -0,072 | |
| | 0,993 | 0,043 | 1,046 | 0,062 | 0,054 | <i>Sylvia atricapilla</i> | 1,038 | 0,016 | 0,970 | 0,018 | -0,068 | |
| | 1,195 | 0,017 | 1,112 | 0,018 | -0,084 | <i>Phylloscopus bonelli</i> | 1,001 | 0,010 | 0,933 | 0,011 | -0,068 | |
| | 1,323 | 0,028 | 1,319 | 0,034 | -0,003 | <i>Phylloscopus collybita</i> | 1,097 | 0,034 | 0,953 | 0,031 | -0,144 | |
| | 1,013 | 0,007 | 0,957 | 0,008 | -0,057 | <i>Regulus ignicapillus</i> | 1,204 | 0,056 | 1,352 | 0,104 | 0,147 | |
| | 1,000 | 0,007 | 0,943 | 0,008 | -0,057 | <i>Muscicapa striata</i> | 1,011 | 0,019 | 0,933 | 0,021 | -0,078 | |
| | 1,096 | 0,012 | 1,114 | 0,015 | 0,018 | <i>Aegithalos caudatus</i> | 0,996 | 0,041 | 0,874 | 0,041 | -0,122 | |
| | 1,166 | 0,058 | 1,226 | 0,079 | 0,060 | <i>Parus palustris</i> | 0,940 | 0,031 | 0,908 | 0,038 | -0,032 | |
| | 1,053 | 0,014 | 1,041 | 0,017 | -0,012 | <i>Parus ater</i> | 1,127 | 0,011 | 1,036 | 0,012 | -0,091 | |
| | 0,960 | 0,026 | 0,892 | 0,030 | -0,069 | <i>Parus caeruleus</i> | 1,143 | 0,010 | 1,103 | 0,011 | -0,040 | |
| | 0,977 | 0,013 | 0,830 | 0,014 | -0,147 | <i>Parus major</i> | 1,009 | 0,025 | 0,953 | 0,028 | -0,057 | |
| | 1,066 | 0,037 | 0,972 | 0,039 | -0,094 | <i>Sitta europaea</i> | 1,062 | 0,025 | 1,000 | 0,028 | -0,062 | |
| | 1,213 | 0,029 | 1,241 | 0,040 | 0,028 | <i>Certhia brachydactyla</i> | 1,087 | 0,014 | 1,034 | 0,016 | -0,054 | |
| | 1,064 | 0,015 | 0,984 | 0,016 | -0,080 | <i>Oriolus oriolus</i> | 0,996 | 0,018 | 1,001 | 0,024 | 0,005 | |
| | 0,987 | 0,011 | 0,948 | 0,013 | -0,039 | <i>Lanius collurio</i> | 0,988 | 0,012 | 0,916 | 0,013 | -0,072 | |
| | 1,040 | 0,009 | 1,059 | 0,012 | 0,019 | <i>Garrulus glandarius</i> | 1,075 | 0,016 | 1,050 | 0,019 | -0,025 | |
| | 1,075 | 0,012 | 1,062 | 0,016 | -0,013 | <i>Pica pica</i> | 1,054 | 0,027 | 1,059 | 0,036 | 0,004 | |
| | 1,154 | 0,054 | 1,207 | 0,076 | 0,053 | <i>Corvus monedula</i> | 0,993 | 0,006 | 0,961 | 0,007 | -0,031 | |
| | 1,019 | 0,019 | 0,988 | 0,022 | -0,031 | <i>Corvus corone cornix</i> | 1,075 | 0,012 | 1,046 | 0,014 | -0,029 | |
| | 0,987 | 0,009 | 0,916 | 0,010 | -0,072 | <i>Sturnus vulgaris</i> | 0,983 | 0,008 | 0,990 | 0,010 | 0,007 | |
| | 1,068 | 0,008 | 0,977 | 0,008 | -0,091 | <i>Passer italiae</i> | 1,120 | 0,020 | 1,074 | 0,023 | -0,046 | |
| | 1,008 | 0,007 | 0,952 | 0,009 | -0,057 | <i>Passer montanus</i> | 1,038 | 0,005 | 0,991 | 0,006 | -0,047 | |
| | 1,105 | 0,057 | 1,027 | 0,062 | -0,078 | <i>Fringilla coelebs</i> | 1,001 | 0,006 | 0,966 | 0,008 | -0,036 | |
| | 1,308 | 0,075 | 1,269 | 0,077 | -0,040 | <i>Serinus serinus</i> | 0,979 | 0,009 | 0,927 | 0,011 | -0,052 | |
| | 1,052 | 0,015 | 1,017 | 0,018 | -0,034 | <i>Carduelis chloris</i> | 0,973 | 0,007 | 0,932 | 0,008 | -0,041 | |
| | 0,945 | 0,037 | 0,923 | 0,045 | -0,022 | <i>Carduelis carduelis</i> | 1,052 | 0,025 | 0,943 | 0,026 | -0,109 | |
| | 1,082 | 0,005 | 1,006 | 0,005 | -0,075 | <i>Carduelis cannabina</i> | 0,893 | 0,055 | 0,737 | 0,054 | -0,156 | |
| | 1,118 | 0,048 | 1,036 | 0,051 | -0,082 | <i>Emberiza citrinella</i> | 0,971 | 0,008 | 0,919 | 0,009 | -0,052 | |
| | 0,960 | 0,017 | 0,890 | 0,019 | -0,070 | <i>Emberiza cirrus</i> | 1,036 | 0,049 | 1,085 | 0,069 | 0,048 | |
| | 0,963 | 0,013 | 0,936 | 0,016 | -0,028 | <i>Emberiza cia</i> | 1,000 | 0,011 | 0,970 | 0,014 | -0,030 | |
| | 1,011 | 0,022 | 1,034 | 0,029 | 0,023 | <i>Miliaria calandra</i> | | | | | | |

crease between the first and second years followed by a recovery. Differences between FBI and WBI trends are even more pronounced if the values obtained within the habitat of interest are considered.

DISCUSSION

The data collected up to now seem to show that in Umbria

a great part of common bird species is declining. As a matter of fact, 42 % of the species whose trend has been produced have decreased over the period of the study, whilst only 28 % have increased. The situation appears worst for woodland species compared with farmland ones, as shown by the trends of WBI and FBI. This result is quite different from the general European findings, that show a relative stability of WBI and a sharp decline of FBI (Gregory *et al.*, op.cit.). However, our results must be considered on-

ly as preliminary, since they are based on a small number of years.

From a methodological point of view, we have verified that the trends of FBI and WBI remain basically unchanged both using all the sites (points) available, or previously selecting them on the basis of their environmental features. This result corresponds to the conclusions who got Newson *et al.* (2004) based on experiences made in the UK and suggests that in wide monitoring scheme reliable

FBI and WBI calculations are possible even if the habitat characteristics in the surveyed sites are not recorded.

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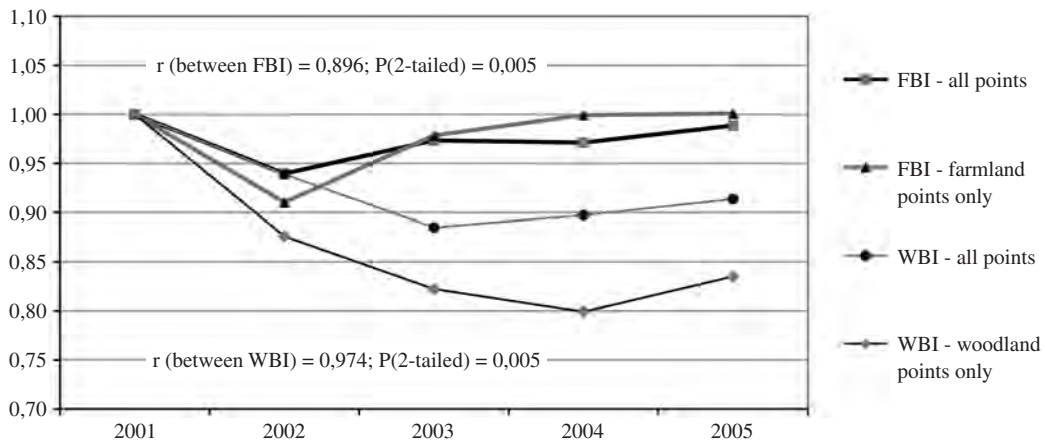


Figure 1. Trends of the multi-species indicators over 2001-2005.

Table 3. Results of the trend estimation for common breeding birds in Umbria, Italy for the period 2001-2005. We present the trend estimate (b) with standard errors and 95 % confidence intervals. Trends are classified as stable (S), moderate decline (MD), steep decline (SD), moderate increase (MI), strong increase (SI), and trend uncertain (U), according to the criteria shown in Table 1. F = farmland; W = woodland.

| SPECIES | habitat | number of points of occurrence | b | SE | 95% confidence interval | | TREND |
|------------------------------|---------|--------------------------------|-------|-------|-------------------------|-------------|-------|
| | | | | | lower limit | upper limit | |
| <i>Buteo buteo</i> | F | 366 | 0,929 | 0,026 | 0,877 | 0,980 | MD |
| <i>Falco tinnunculus</i> | F | 254 | 1,075 | 0,039 | 0,999 | 1,150 | U |
| <i>Coturnix coturnix</i> | | 229 | 1,158 | 0,031 | 1,097 | 1,218 | SI |
| <i>Phasianus colchicus</i> | | 760 | 0,953 | 0,012 | 0,929 | 0,978 | MD |
| <i>Gallinula chloropus</i> | | 39 | 1,046 | 0,062 | 0,925 | 1,167 | U |
| <i>Columba palumbus</i> | | 747 | 1,112 | 0,018 | 1,077 | 1,147 | SI |
| <i>Streptopelia decaocto</i> | | 280 | 1,319 | 0,034 | 1,253 | 1,385 | SI |
| <i>Streptopelia turtur</i> | F | 1234 | 0,957 | 0,008 | 0,941 | 0,973 | MD |
| <i>Cuculus canorus</i> | | 1258 | 0,943 | 0,008 | 0,928 | 0,959 | MD |
| <i>Apus apus</i> | | 1050 | 1,114 | 0,015 | 1,085 | 1,143 | SI |
| <i>Merops apiaster</i> | | 68 | 1,226 | 0,079 | 1,070 | 1,381 | SI |
| <i>Upupa epops</i> | F | 746 | 1,041 | 0,017 | 1,007 | 1,074 | MI |
| <i>Jynx torquilla</i> | | 287 | 0,892 | 0,030 | 0,833 | 0,950 | MD |
| <i>Picus viridis</i> | W | 788 | 0,830 | 0,014 | 0,802 | 0,857 | SD |
| <i>Picoides major</i> | W | 194 | 0,972 | 0,039 | 0,895 | 1,049 | U |
| <i>Galerida cristata</i> | F | 217 | 1,241 | 0,040 | 1,162 | 1,319 | SI |



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| SPECIES | habitat | number of points of occurrence | b | SE | 95% confidence interval | | TREND |
|--------------------------------|---------|--------------------------------|-------|-------|-------------------------|-------------|-------|
| | | | | | lower limit | upper limit | |
| <i>Lullula arborea</i> | | 584 | 0,984 | 0,016 | 0,952 | 1,015 | S |
| <i>Alauda arvensis</i> | F | 415 | 0,948 | 0,013 | 0,923 | 0,974 | MD |
| <i>Hirundo rustica</i> | F | 1009 | 1,059 | 0,012 | 1,034 | 1,083 | MI |
| <i>Delichon urbica</i> | F | 746 | 1,062 | 0,016 | 1,031 | 1,092 | MI |
| <i>Motacilla flava</i> | F | 58 | 1,207 | 0,076 | 1,058 | 1,356 | SI |
| <i>Motacilla alba</i> | F | 465 | 0,988 | 0,022 | 0,944 | 1,032 | U |
| <i>Troglodytes troglodytes</i> | W | 935 | 0,916 | 0,010 | 0,896 | 0,936 | SD |
| <i>Erithacus rubecula</i> | W | 1073 | 0,977 | 0,008 | 0,961 | 0,994 | MD |
| <i>Luscinia megarhynchos</i> | F | 1058 | 0,952 | 0,009 | 0,935 | 0,968 | MD |
| <i>Phoenicurus ochruros</i> | | 65 | 1,027 | 0,062 | 0,905 | 1,148 | U |
| <i>Phoenicurus phoenicurus</i> | | 75 | 1,269 | 0,077 | 1,117 | 1,420 | SI |
| <i>Saxicola torquata</i> | F | 607 | 1,017 | 0,018 | 0,983 | 1,052 | U |
| <i>Oenanthe oenanthe</i> | | 37 | 0,923 | 0,045 | 0,834 | 1,012 | U |
| <i>Turdus merula</i> | | 1541 | 1,006 | 0,005 | 0,996 | 1,016 | S |
| <i>Turdus viscivorus</i> | W | 126 | 1,036 | 0,051 | 0,936 | 1,136 | U |
| <i>Cettia cetti</i> | F | 237 | 0,890 | 0,019 | 0,852 | 0,927 | SD |
| <i>Cisticola juncidis</i> | F | 429 | 0,936 | 0,016 | 0,905 | 0,967 | MD |
| <i>Hippolais polyglotta</i> | F | 277 | 1,034 | 0,029 | 0,977 | 1,091 | U |
| <i>Sylvia cantillans</i> | W | 512 | 1,064 | 0,021 | 1,024 | 1,105 | MI |
| <i>Sylvia melanocephala</i> | | 414 | 0,937 | 0,019 | 0,899 | 0,974 | MD |
| <i>Sylvia communis</i> | | 229 | 0,858 | 0,030 | 0,800 | 0,916 | SD |
| <i>Sylvia atricapilla</i> | W | 1549 | 0,984 | 0,005 | 0,975 | 0,994 | MD |
| <i>Phylloscopus bonelli</i> | W | 341 | 0,970 | 0,018 | 0,935 | 1,005 | U |
| <i>Phylloscopus collybita</i> | W | 702 | 0,933 | 0,011 | 0,911 | 0,955 | MD |
| <i>Regulus ignicapillus</i> | W | 204 | 0,953 | 0,031 | 0,893 | 1,013 | U |
| <i>Muscicapa striata</i> | W | 121 | 1,352 | 0,104 | 1,148 | 1,555 | SI |
| <i>Aegithalos caudatus</i> | W | 564 | 0,933 | 0,021 | 0,893 | 0,974 | MD |
| <i>Parus palustris</i> | W | 129 | 0,874 | 0,041 | 0,794 | 0,954 | MD |
| <i>Parus ater</i> | W | 100 | 0,908 | 0,038 | 0,833 | 0,982 | MD |
| <i>Parus caeruleus</i> | W | 1110 | 1,036 | 0,012 | 1,013 | 1,058 | MI |
| <i>Parus major</i> | | 1292 | 1,103 | 0,011 | 1,081 | 1,124 | SI |
| <i>Sitta europaea</i> | W | 249 | 0,953 | 0,028 | 0,898 | 1,008 | U |
| <i>Certhia brachydactyla</i> | W | 310 | 1,000 | 0,028 | 0,946 | 1,054 | U |
| <i>Oriolus oriolus</i> | | 755 | 1,034 | 0,016 | 1,003 | 1,064 | MI |
| <i>Lanius collurio</i> | F | 394 | 1,001 | 0,024 | 0,954 | 1,048 | S |
| <i>Garrulus glandarius</i> | W | 817 | 0,916 | 0,013 | 0,890 | 0,942 | SD |
| <i>Pica pica</i> | F | 415 | 1,050 | 0,019 | 1,013 | 1,087 | MI |
| <i>Corvus monedula</i> | | 273 | 1,059 | 0,036 | 0,988 | 1,129 | U |
| <i>Corvus corone cornix</i> | F | 1521 | 0,961 | 0,007 | 0,948 | 0,975 | MD |
| <i>Sturnus vulgaris</i> | F | 922 | 1,046 | 0,014 | 1,018 | 1,074 | MI |
| <i>Passer italiae</i> | F | 914 | 0,990 | 0,010 | 0,970 | 1,009 | S |
| <i>Passer montanus</i> | F | 467 | 1,074 | 0,023 | 1,030 | 1,119 | MI |
| <i>Fringilla coelebs</i> | W | 1410 | 0,991 | 0,006 | 0,980 | 1,003 | S |
| <i>Serinus serinus</i> | F | 1176 | 0,966 | 0,008 | 0,951 | 0,980 | MD |
| <i>Carduelis chloris</i> | F | 1020 | 0,927 | 0,011 | 0,905 | 0,948 | SD |
| <i>Carduelis carduelis</i> | F | 1266 | 0,932 | 0,008 | 0,916 | 0,948 | SD |
| <i>Carduelis cannabina</i> | F | 208 | 0,943 | 0,026 | 0,892 | 0,994 | MD |
| <i>Emberiza citrinella</i> | | 55 | 0,737 | 0,054 | 0,631 | 0,844 | SD |
| <i>Emberiza cirlus</i> | F | 1046 | 0,919 | 0,009 | 0,901 | 0,937 | SD |
| <i>Emberiza cia</i> | | 58 | 1,085 | 0,069 | 0,950 | 1,219 | U |
| <i>Miliaria calandra</i> | F | 498 | 0,970 | 0,014 | 0,943 | 0,996 | MD |

Table 4. Farmland species: comparison between the analysis carried out on the whole set of points of occurrence and only on the points included in farmland habitat. We present the trend assessments under the two methods and the Pearson correlations between population indices (* correlation significant at the 5 % level; ** correlation significant at the 1 % level). Trend abbreviations as in Table 3.

| SPECIES | number of points of occurrence | | points in farmland habitats (%) | TREND | | Pearson correlation (N=5) between population indices obtained in farmland habitat and in the whole set of points | |
|------------------------------|--------------------------------|--------------|---------------------------------|------------------|--------------|--|--------------|
| | farmland habitat | all habitats | | farmland habitat | all habitats | r | P (2-tailed) |
| | | | | | | | |
| <i>Falco tinnunculus</i> | 150 | 254 | 59,1 | U | U | 0,986 | 0,002 ** |
| <i>Streptopelia turtur</i> | 694 | 1234 | 56,2 | MD | MD | 0,927 | 0,023 * |
| <i>Upupa epops</i> | 407 | 746 | 54,6 | U | MI | 0,838 | 0,076 |
| <i>Galerida cristata</i> | 206 | 217 | 94,9 | SI | SI | 0,999 | 0,000 ** |
| <i>Alauda arvensis</i> | 296 | 415 | 71,3 | MD | MD | 0,940 | 0,018 * |
| <i>Hirundo rustica</i> | 694 | 1009 | 68,8 | MI | MI | 0,975 | 0,005 ** |
| <i>Delichon urbica</i> | 500 | 746 | 67,0 | MI | MI | 0,912 | 0,031 * |
| <i>Motacilla flava</i> | 58 | 58 | 100,0 | SI | SI | 1,000 | 0,000 ** |
| <i>Motacilla alba</i> | 282 | 465 | 60,6 | U | U | 0,855 | 0,065 |
| <i>Luscinia megarhynchos</i> | 709 | 1058 | 67,0 | MD | MD | 0,974 | 0,005 ** |
| <i>Saxicola torquata</i> | 445 | 607 | 73,3 | U | U | 0,940 | 0,017 * |
| <i>Cettia cetti</i> | 195 | 237 | 82,3 | SD | SD | 0,993 | 0,001 ** |
| <i>Cisticola juncidis</i> | 371 | 429 | 86,5 | MD | MD | 0,999 | 0,000 ** |
| <i>Hippolais polyglotta</i> | 193 | 277 | 69,7 | U | U | 0,978 | 0,004 ** |
| <i>Lanius collurio</i> | 238 | 394 | 60,4 | U | S | 0,762 | 0,134 |
| <i>Pica pica</i> | 316 | 415 | 76,1 | MI | MI | 0,948 | 0,014 * |
| <i>Corvus corone cornix</i> | 810 | 1521 | 53,3 | MD | MD | 0,910 | 0,032 * |
| <i>Sturnus vulgaris</i> | 649 | 922 | 70,4 | MI | MI | 0,782 | 0,118 |
| <i>Passer italiae</i> | 674 | 914 | 73,7 | S | S | 0,986 | 0,002 ** |
| <i>Passer montanus</i> | 444 | 467 | 95,1 | MI | MI | 0,985 | 0,002 ** |
| <i>Serinus serinus</i> | 757 | 1176 | 64,4 | MD | MD | 0,964 | 0,008 ** |
| <i>Carduelis chloris</i> | 635 | 1020 | 62,3 | MD | SD | 0,935 | 0,020 * |
| <i>Carduelis carduelis</i> | 779 | 1266 | 61,5 | MD | SD | 0,985 | 0,002 ** |
| <i>Carduelis cannabina</i> | 74 | 208 | 35,6 | U | MD | 0,721 | 0,169 |
| <i>Emberiza cirulus</i> | 496 | 1046 | 47,4 | MD | SD | 0,931 | 0,022 * |
| <i>Miliaria calandra</i> | 326 | 498 | 65,5 | S | MD | 0,742 | 0,151 |

Table 5. Woodland species: comparison between the analysis carried out on the whole set of points of occurrence and only on the points included in woodland habitat. We present the trend assessments under the two methods and the Pearson correlations between population indices (* correlation significant at the 5 % level; ** correlation significant at the 1 % level). Trend abbreviations as in Table 3.

| SPECIES | number of points of occurrence | | points in woodland habitats (%) | TREND | | Pearson correlation (N=5) between population indices obtained in woodland habitat and in the whole set of points | |
|--------------------------------|--------------------------------|--------------|---------------------------------|------------------|--------------|--|--------------|
| | woodland habitat | all habitats | | woodland habitat | all habitats | r | P (2-tailed) |
| | | | | | | | |
| <i>Picoides major</i> | 86 | 194 | 44,3 | U | U | 0,108 | 0,863 |
| <i>Troglodytes troglodytes</i> | 362 | 935 | 38,7 | SD | SD | 0,988 | 0,002 ** |
| <i>Erithacus rubecula</i> | 449 | 1073 | 41,8 | MD | MD | 0,961 | 0,009 ** |
| <i>Turdus viscivorus</i> | 69 | 126 | 54,8 | U | U | -0,214 | 0,730 |
| <i>Sylvia cantillans</i> | 178 | 512 | 34,8 | U | MI | 0,442 | 0,456 |
| <i>Sylvia atricapilla</i> | 459 | 1549 | 29,6 | MD | MD | 0,967 | 0,007 ** |



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| SPECIES | number of points of occurrence | | points in woodland habitats (%) | TREND | | Pearson correlation (N=5) between population indices obtained in woodland habitat and in the whole set of points | |
|-------------------------------|--------------------------------|--------------|---------------------------------|------------------|--------------|--|--------------|
| | woodland habitat | all habitats | | woodland habitat | all habitats | r | P (2-tailed) |
| | | | | | | | |
| <i>Phylloscopus collybita</i> | 357 | 702 | 50,9 | MD | MD | 0,992 | 0,001 ** |
| <i>Regulus ignicapillus</i> | 114 | 204 | 55,9 | MD | U | 0,929 | 0,022 * |
| <i>Muscicapa striata</i> | 24 | 121 | 19,8 | U | SI | 0,507 | 0,383 |
| <i>Aegithalos caudatus</i> | 214 | 564 | 37,9 | MD | MD | 0,903 | 0,036 * |
| <i>Parus palustris</i> | 83 | 129 | 64,3 | MD | MD | 0,965 | 0,008 ** |
| <i>Parus ater</i> | 74 | 100 | 74,0 | MD | MD | 0,971 | 0,006 ** |
| <i>Parus caeruleus</i> | 394 | 1110 | 35,5 | U | MI | 0,962 | 0,009 ** |
| <i>Sitta europaea</i> | 144 | 249 | 57,8 | U | U | 0,918 | 0,028 * |
| <i>Certhia brachydactyla</i> | 119 | 310 | 38,4 | U | U | -0,476 | 0,418 |
| <i>Garrulus glandarius</i> | 375 | 817 | 45,9 | SD | SD | 0,944 | 0,016 * |
| <i>Fringilla coelebs</i> | 450 | 1410 | 31,9 | S | S | 0,975 | 0,005 ** |

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